Prenatal Exposure to Phthalates and Childhood Body Size in an Urban Cohort

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BACKGROUND: Phthalate exposures are hypothesized to increase obesity; however, prior research has been largely cross-sectional.

OBJECTIVE: We evaluated associations between prenatal phthalate exposures and body mass index (BMI) at child ages 5 and 7 years.

METHODS: Nine metabolites of six phthalates—di(2-ethylhexyl) phthalate (DEHP), di-n-octyl-, di-iso-butyl-, di-n-butyl-, butylbenzyl-, and diethyl phthalates—were measured in spot urine samples collected from pregnant African-American and Dominican women during their third trimester, and from their children at ages 3 and 5 years. To reduce multiple comparison issues, we initially used principal component analysis (PCA) to identify major patterns of natural log (ln)-transformed metabolite concentrations. Height and weight were assessed at ages 5 and 7 years, and fat mass and waist circumference at age 7. Linearized generalized estimating equation analyses related maternal component scores to child anthropometric outcomes at ages 5 (n = 326) and 7 (n = 330) years.

RESULTS: PCA identified a DEHP component and a non-DEHP component. In boys, higher maternal non-DEHP, but not DEHP, component scores were associated with lower BMI z-score ($\beta = -0.30$; 95% CI: -0.50, -0.10, n = 156), lower fat percentage ($\beta = -1.62$; 95% CI: -2.91, -0.34, n = 142), and smaller waist circumference ($\beta = -2.02$; 95% CI: -3.71, -0.32, n = 124). No significant associations with anthropometric outcomes were seen in girls (for BMI z-score, $\beta = 0.07$; 95% CI: -0.18, 0.31, n = 181). Interactions between sex and non-DEHP component association with outcomes were statistically significant (p < 0.01).

Conclusions: Contrary to hypotheses, prenatal non-DEHP phthalate exposures were associated with lower BMI z-score, waist circumference, and fat mass in boys during early childhood.

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Introduction

Certain endocrine-disrupting chemicals (EDCs), including phthalates, are hypothesized to be obesogens and contribute to the obesity epidemic (Grün and Blumberg 2009; Heindel 2003; Newbold et al. 2008). Phthalates are a class of EDCs with ubiquitous human exposure, commonly found in plastics, fragrances, cosmetics, and building materials (Sathyanarayana 2008; Silva et al. 2004). Phthalates have anti-androgenic and weakly estrogenic properties, and have been found to increase adipogenesis in vitro and in some animal studies (Biemann et al. 2012; Hao et al. 2012; Harris et al. 1997; Jobling et al. 1995; Lovekamp and Davis 2001; Rengarajan et al. 2007; Takeuchi et al. 2005). However, other animal studies have shown either a decrease or no significant change in adipose weight after prenatal phthalate exposure (Boberg et al. 2008; Casals-Casas et al. 2008; Casals-Casas and Desvergne 2011). Crosssectional studies using National Health and Nutrition Examination Survey (NHANES) data found that urinary concentrations of five phthalate metabolites [monobenzyl phthalate (MBzP), mono(2-ethyl-5-oxohexyl) phthalate

(MEOHP), mono(2-ethyl-5-hydroxyhexyl) phthalate (MEHHP), monoethyl phthalate (MEP), mono-n-butyl phthalate (MBP)] correlated positively with body mass index (BMI) in adult males (Hatch et al. 2008; Stahlhut et al. 2007), and similarly MEP correlated positively with BMI in adolescent females (Hatch et al. 2008). However, an inverse correlation was observed between mono(2-ethylhexyl) phthalate (MEHP) and BMI in adolescent and adult females (Hatch et al. 2008). In cross-sectional analyses among children, lowmolecular-weight phthalates were positively associated with BMI and odds of overweight and obesity, but only among non-Hispanic blacks (Trasande et al. 2013a). A study by Teitelbaum et al. (2012) found that higher urinary concentrations of MEP and molar sum of MEP, MBP, and mono-isobutyl phthalate (MiBP) measured at age 6-8 years were associated with higher BMI a year later among overweight children. However, previous studies have been limited by cross-sectional designs or short follow-up periods, and have not evaluated potential effects of exposures during critical developmental periods of fetal and early newborn life.

Using the Columbia Center for Children's Environmental Health (CCCEH) longitudinal birth cohort, we conducted prospective analyses of maternal urinary phthalate metabolite concentrations and child body size at ages 5 and 7 years, with further assessment of associations between body size and childhood phthalate exposures at ages 3 and 5 years. We hypothesized that prenatal phthalate exposures would be associated with higher BMI z-scores (age- and sex-adjusted

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A.G.R. is on the Medical Advisory Board of EHE International, Inc., which provides wellness programs and annual physical exams to employees of large corporations. He provides expertise on the obesity epidemic, development of programs to combat obesity, and analyses of medical data from their patients, and receives \$10,000 per year in compensation. This work does not overlap with his role as a co-author in the submitted manuscript. The other authors declare they have no competing financial interests.

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BMI), greater fat mass, and larger waist circumferences and that associations would be stronger among boys due to phthalates' possible anti-androgenic effect.

Methods

Participant selection and data collection. The children in this study are enrolled in the CCCEH longitudinal birth cohort study's Obesity Project, which has been described extensively elsewhere (Perera et al. 2003; Rundle et al. 2012). Recruitment took place between 1998 and 2006 through prenatal clinics in northern Manhattan, at New York-Presbyterian Hospital and Harlem Hospital Center. Inclusion criteria were pregnant women between 18 and 35 years of age, who self-identified as either African American or Dominican, and resided in Northern Manhattan or the South Bronx in New York City for at least 1 year before pregnancy. Exclusion criteria were current smoking at time of recruitment, illicit drug use, or having a diagnosis of HIV, hypertension, or diabetes. During the third trimester of pregnancy, a trained bilingual interviewer collected sociodemographic information from participants. Infants' sex and birth weight were obtained from medical records following delivery. As described previously, at ages 5 and 7 years the children's heights were measured on a SECA wall-mounted stadiometer to the nearest 0.1 cm (SECA), age 5 weight was measured to the nearest 0.1 kg using a Detecto Cardinal 750 digital scale (Cardinal Scale Manufacturing Company), and at age 7 years total weight and fat mass were measured using a Tanita scale (Model BC-418; Tanita Corporation of America). Measurements were taken while wearing light clothes and no shoes (Rundle et al. 2012). BMI z-scores and percentiles were calculated using the Centers for Disease Control and Prevention (CDC) SAS macro; obesity was defined as BMI for age and sex > 95th percentile (CDC 2004). Waist circumference was measured midway between the lower rib margin and iliac crest using a nonstretching tape measure (Rundle et al. 2012).

The institutional review boards of Columbia University Medical Center and the CDC approved this study. All mothers gave informed consent at enrollment, and all children at age 7 years assented to participate.

Measurement of urinary phthalate metabolites. Spot urine samples were collected from the mothers during the third trimester of pregnancy, and from the children at ages 3 and 5 years using a urine cup supplied by the CDC. The specific gravity of the urine samples was quantified at room temperature at Columbia University with a handheld refractometer (PAL 10-S; Atago). Samples were frozen at -80°C immediately after collection

and shipped frozen to the CDC for analysis. Using solid phase extraction-high performance liquid chromatography/isotope dilution tandem mass spectrometry (Silva et al. 2007), urine samples were analyzed for metabolites of di(2-ethylhexyl) phthalate (DEHP), which are MEHP, MEHHP, MEOHP, and mono(2-ethyl-5-carboxypentyl) phthalate (MECPP); and for metabolites of di-isobutyl phthalate (DiBP), di-n-butyl phthalate (DnBP), butylbenzyl phthalate (BBzP), diethyl phthalate (DEP), and di-n-octyl phthalate (DOP), which are, respectively, MiBP, MBP, MBzP, and MEP, and mono(3-carboxypropyl) phthalate (MCPP), a nonspecific metabolite of several phthalates.

Data analysis. All analyses were performed using PASW® Statistics 21 (SPSS®) and SAS version 9.3 (SAS institute Inc.). The value of the limit of detection (LOD)/2 was assigned to phthalate metabolite concentrations below the LOD [MEHP: 54 samples (LOD, 0.5-1.2 ng/dL), MCPP: 18 samples (LOD, 0.16-1 ng/dL), MiBP (LOD, 0.2-1.038 ng/dL), and MBzP (LOD, 0.11-1 ng/dL); 1 sample each]. As recommended by the CDC, reported concentrations of MEP and MBzP were multiplied by 0.66 and 0.72, respectively, to correct for impurity in the analytic standards which led to overestimation of the concentrations of these metabolites (CDC 2014). Urinary phthalate metabolite concentrations were natural log (ln)-transformed for analyses.

Correlations between In-transformed concentrations of urinary metabolites were assessed using Pearson product moment correlation analyses (see Supplemental Material, Table S1). Because of medium to strong correlations between urinary phthalate metabolite concentrations (0.34-0.99, p < 0.001 for all correlations), and to reduce multiple comparisons, the initial statistical analytical plan used principal component analysis (PCA) to generate summary measures of phthalate exposure patterns from the urinary metabolite data (Hastie et al. 2009; Kleinbaum et al. 1988). PCA with varimax rotation was applied to In-transformed maternal and child (at ages 3 and 5 years) urinary metabolite concentrations. PCA is a data reduction tool that describes variability among multiple correlated, observed variables in terms of a fewer number of unobserved variables, referred to as components, and has been used to identify patterns of exposure from complex chemical mixtures (Arif and Shah 2007; Bengraïne and Marhaba 2003; Bergh et al. 2011; Billionnet et al. 2012; Burstyn 2004; Dominici et al. 2010; Guo et al. 2004; Lampa et al. 2012; Zitko 1994). Component scores were calculated for the mothers and the children at ages 3 and 5 years, and represent how closely the metabolites in a subject's urine sample conform to identified metabolite patterns. Because PCA results depend on the correlation structure within a particular data set, the correlation and PCA analyses were also applied to 2003-2004 NHANES phthalate urinary metabolite data to ascertain whether the components identified in this study data were consistent with nationally representative data (CDC 2014). Phthalate urinary metabolite concentration data for MEHP, MEHHP, MECPP, MEOHP, MCPP, MiBP, MBP, MBzP, and MEP from women ages 15-45 years who took part in the 2003-2004 NHANES were downloaded (n = 595). This time period best overlaps with enrollment and urine collection in the CCCEH. The metabolite data were In-transformed and analyzed using PCA with varimax rotation. Analyses incorporated the NHANES complex survey weights.

The molar sum of DEHP metabolite (MEHP, MEHHP, MEOHP, and MECPP) concentrations in maternal urine was also calculated.

We used a generalized estimating equation (GEE) model with a linear link function and unstructured correlation structure (Hubbard et al. 2010) to assess whether maternal urinary phthalate concentration component scores were associated with child BMI z-scores at ages 5 and 7 years. We used linear regression analyses to test whether maternal urinary phthalate concentration component scores were associated with child age 7 years percent body fat, fat mass index (FMI; fat mass in kilograms/height in meters squared) and waist circumference (centimeters). All models were adjusted for maternal race/ethnicity (selfreported as African American or Dominican), receipt of public assistance during pregnancy, prepregnancy obesity status (based on mother's self-reported height and prepregnancy weight), child birth weight (continuous variable), child age in months at the time of follow-up at ages 5 and 7 years, and maternal urine specific gravity z-score to adjust for urine concentration because it is associated with component scores (though less strongly associated with raw metabolite concentrations). Analyses were conducted separately in boys and girls, and a model incorporating sex by component score interaction terms was fit to formally test whether associations varied by sex. Secondary analyses were undertaken when significant associations (p < 0.05) were identified in stratified analyses, assessing relationships between the anthropometric outcome and individual phthalate metabolites loading heavily for the component.

A series of sensitivity analyses were conducted to assess potential confounding effects by phthalate exposures at child age 3 and 5 years, stability of the results to different model specifications, and possible bias from loss to follow-up. To assess possible confounding between maternal and child phthalate exposures, we assessed correlations between the mothers' component scores and the children's ages 3 and 5 years component scores, and the children's component scores were added to the GEE model described above. Possible interactions between maternal urinary phthalate component scores and age 5 versus age 7 follow-up wave were assessed by including a follow-up wave by component scores interaction term in the model. To assess possible bias due to incomplete followup at ages 5 and 7 years, inverse probability weights for successful assessment at each of these two waves were calculated as described previously (Curtis et al. 2007; Hernán et al. 2004; Rundle et al. 2012), and these weights were incorporated into the GEE models.

Results

Prenatal urinary phthalate metabolite concentrations were available from 424 mothers. Geometric mean concentrations of phthalate metabolites in maternal urine samples and intercorrelations between concentrations are reported in Supplemental Material, Table S1. The correlation matrix observed for the mothers' metabolite concentrations was very similar to that observed in NHANES (see Supplemental Material, Table S2). Cohort follow-up and urinary metabolite data availability are documented in Supplemental Material, Figure S1. From the 424 mothers with urinary phthalate data, 326 children had anthropometric outcome data at age 5 years (310 with complete covariate data), 330 had anthropometric outcome data at age 7 years (314 with complete covariate data), 303 had outcome data at both age 5 and 7 years, and 353 had outcome data at age 5 or 7 years. Among the 326 children followed to age 5, urinary metabolite data were available from 234 children at age 3 years. Among the 330 children followed to age 7, urinary metabolite data were available from 241 at age 3 and 302 at age 5 years.

Table 1 provides descriptive data for the cohort. Nineteen percent of the children were obese at age 5 and 25% were obese at age 7 years. PCA identified two components in the maternal urinary metabolite data with eigenvalues > 1 (see Supplemental Material, Figure S2), one component in the child age 3 data, and two components in the child age 5 data, representing major patterns of phthalate metabolite concentrations among participants (Table 2). The first maternal component, labeled the "DEHP" component, accounted for 61% of the variation in the data; it loaded highly for DEHP metabolites and was highly correlated with the molar sum of DEHP metabolites (r = 0.95). The second component,

the "non-DEHP" component, accounted for an additional 16% of the variation in the data and loaded highly for non-DEHP metabolites. PCA analysis of child age 5 phthalate metabolite concentrations also identified a DEHP and non-DEHP component, but analysis of child age 3 urine revealed only a single component. PCA of NHANES phthalate metabolite data identified two components with very similar loadings to the components observed in the CCCEH mothers' data (see Supplemental Material, Table S3).

Table 3 displays associations between maternal phthalate metabolite component

scores and child BMI z-scores at ages 5 and 7 years. Prenatal DEHP component scores were not significantly associated with BMI z-scores at ages 5 and 7. In all three models, prenatal non-DEHP component score was significantly associated with lower BMI z-scores in boys only. Associations of child BMI z-scores with prenatal DEHP and non-DEHP component scores were similar after adjustment for component scores from the children's urine samples collected at age 3 (model 2) or at age 5 years (model 3). In model 1 (which did not include the children's urine metabolite concentrations), the

Table 1. Characteristics of mothers and children with urinary phthalate measurements and childhood anthropometric outcomes [n(%)] or mean \pm SD unless otherwise specified].

	Total enrolled With age 5		With age 7	
	cohort	anthropometric data	anthropometric data	
Characteristic	(n = 424)	(n = 326)	(n = 330)	
Sex of child				
Female	219 (52)	178 (55)	173 (52)	
Male	205 (48)	148 (45)	157 (48)	
Ethnicity				
African American	139 (33)	119 (36.5)	119 (36)	
Dominican	285 (67)	207 (63.5)	211 (64)	
Maternal prepregnancy obesity (BMI > 30)				
Yes	85 (20)	70 (22)	71 (22)	
No	328 (77)	248 (76)	251 (76)	
Missing	11 (3)	8 (2.5)	8 (2)	
Maternal receipt of public assistance				
Yes	182 (56)	141 (43)	140 (42)	
No	238 (43)	183 (56)	188 (57)	
Missing	4 (1)	2 (1)	2 (1)	
Child obesity ^a				
Yes	NA	63 (19)	82 (25)	
No	NA	263 (79)	248 (75)	
Mean birth weight (g)	3,375 (± 472)	3,372 (± 481)	3,383 (± 489)	
Mean BMI z-score	NA	0.58 (± 1.43)	0.83 (± 1.16)	
Median BMI percentile (interquartile range)	NA	72.00 (35.44, 92.83)	81.76 (54.09, 95.13)	
Mean percent body fat ($n = 322$)	NA	NA	24.38 (± 6.34)	
Mean fat mass index $(n = 322)$	NA	NA	4.62 (± 2.33)	
Waist circumference (cm) ($n = 318$)	NA	NA	59.91 (± 8.66)	
DEHP component score	0.00 (± 1.00)	0.03 (± 0.99)	0.00 (± 0.99)	
Non-DEHP component score	0.00 (± 1.00)	0.04 (± 0.98)	0.01 (± 1.00)	

Abbreviations: BMI, body mass index; DEHP, di(2-ethylhexyl) phthalate; NA, not applicable.
^aChild obesity defined as BMI ≥ 95th percentile for age and sex according to the CDC growth charts.

Table 2. Rotated component-loading weights for phthalate urinary metabolite concentrations in CCCEH mothers and children.

	Maternal		Child age 3	Child	Child age 5		
Phthalate metabolite	DEHP component ^a	Non-DEHP component ^a	Phthalate component ^b	DEHP component ^c	Non-DEHP component ^c		
MEHP	0.88	0.23	0.81	0.90	0.15		
MEHHP	0.94	0.31	0.95	0.92	0.34		
MECPP	0.92	0.29	0.93	0.90	0.31		
MEOHP	0.92	0.35	0.95	0.91	0.38		
MCPP	0.39	0.68	0.87	0.49	0.60		
MiBP	0.26	0.79	0.83	0.34	0.79		
MBP	0.19	0.89	0.90	0.41	0.83		
MBzP	0.21	0.77	0.78	0.25	0.76		
MEP	0.23	0.60	0.63	0.07	0.66		

Abbreviations: DEHP, di(2-ethylhexyl) phthalate; MBP, mono-n-butyl phthalate; MBzP, monobenzyl phthalate; MCPP, mono(3-carboxypropyl) phthalate; MECPP, mono(2-ethyl-5-carboxypentyl) phthalate; MEHP, mono(2-ethyl-5-hydroxyhexyl) phthalate; MEHP, mono(2-ethylhexyl) phthalate; MEOHP, mono(2-ethyl-5-oxohexyl) phthalate; MEP, monoethyl phthalate; MiBP, mono-isobutyl phthalate.

The DEHP component explains 61% and the non-DEHP component explains 16% of the total variance in the metabolite data. The phthalate component explains 73% of the total variance in the metabolite data. The DEHP component explains 63% and the non-DEHP component explains 14% of the total variance in the metabolite data.

interaction between sex and non-DEHP component score was statistically significant $[\beta = -0.30; 95\% \text{ confidence interval (CI):}$ -0.50, -0.10 in boys compared with $\beta = 0.07$; 95% CI: -0.18, 0.31 in girls; p for interaction = 0.003]. Similarly, higher maternal non-DEHP component scores were significantly associated with lower percent body fat, lower FMI, and smaller waist circumference at age 7 years in boys but not girls (Table 4). For example, for fat percentage, $\beta = -1.62$; 95% CI: -2.91, -0.34 for boys and $\beta = 0.62$; 95% CI: -0.64, 1.88 for girls; p for interaction = 0.002. There were no associations with DEHP component score and percent body fat, FMI, or waist circumference in boys or girls. Secondary analyses examined associations between BMI z-score, fat percentage, FMI, and waist circumference, and individual phthalates that loaded highly for the non-DEHP component. Generally, the associations between the individual metabolites and

anthropometric outcomes did not reach statistical significance, but revealed similar patterns of inverse associations in boys (Table 5). The molar sum of the DEHP metabolites was also analyzed and, as with the DEHP component score, had no associations with BMI *z*-score, fat percentage, FMI, or waist circumference.

Maternal phthalate component scores were not associated with child phthalate component scores (data not shown), and adjustment for child age 3 or age 5 component scores did not alter the associations (Table 3). However, among boys, the child age 5 non-DEHP component score was inversely associated with BMI z-score at ages 5 and 7 years on GEE analysis. Tests of interaction between follow-up wave and maternal urinary metabolite component scores were not statistically significant, and there were no significant differences between the component scores of children who were lost to follow-up and children with follow-up (data not shown).

Discussion

Contrary to our hypothesis that phthalates would be positively associated with body size, our results suggest that prenatal exposures to non-DEHP phthalates, as reflected by higher non-DEHP component scores in maternal urine, are associated with lower BMI z-score, fat mass, and waist circumference in boys at ages 5 and 7 years in our study population. Prenatal exposure to DEHP, as measured by the DEHP component score and molar sum of DEHP metabolites, was not associated with anthropometric outcomes in children through age 7 years. Relationships between maternal urinary phthalate metabolite concentrations and child BMI z-score did not appear to be confounded by the child's urinary phthalate component scores at age 3 years or the DEHP and non-DEHP component scores at age 5 years.

The inverse association between non-DEHP component scores and

Table 3. Associations between maternal prenatal urinary phthalate metabolite component scores and child BMI z-scores at ages 5 and 7 years.

	Model 1 ^a			Model 2 ^b			Model 3 ^c		
Phthalate component score	Girls ($n = 181$) β (95% CI) p-value	Boys (n = 156) β (95% CI) p-value	<i>p</i> -int ^d	Girls (n = 127) β (95% CI) p-value	Boys (n = 113) β (95% CI) p-value	<i>p</i> -int ^d	Girls (n = 154) β (95% CI) p-value	Boys ($n = 124$) β (95% CI) p-value	<i>p</i> -int ^d
Prenatal DEHP component score	-0.10 (-0.29, 0.09) 0.30	-0.08 (-0.30, 0.13) 0.45	NDe	-0.16 (-0.42, 0.09) 0.21	-0.06 (-0.34, 0.22) 0.66	NDe	-0.12 (-0.31, 0.08) 0.25	0.00 (-0.24, 0.24) 0.99	NDe
Prenatal non-DEHP component score	0.07 (-0.18, 0.31) 0.58	-0.30 (-0.50, -0.10) 0.003	0.003	0.16 (-0.17, 0.48) 0.35	-0.31 (-0.56, -0.07) 0.01	0.003	0.13 (-0.14, 0.41) 0.34	-0.30 (-0.54, -0.06) 0.01	0.001
Age 3 phthalate component score	NA	NA	NA	0.02 (-0.34, 0.37) 0.92	0.00 (-0.29, 0.29) 0.99	ND ^e	NA	NA	NA
Age 5 DEHP component score	NA	NA	NA	NA	NA	NA	-0.01 (-0.20, 0.23) 0.91	-0.13 (-0.38, 0.13) 0.33	ND ^e
Age 5 non-DEHP component score	NA	NA	NA	NA	NA	NA	-0.12 (-0.37, 0.14) 0.36	-0.34 (-0.58, -0.10) 0.01	NDe

Abbreviations: DEHP, di-(2-ethylhexyl) phthalate; NA, not applicable; ND, not done.

"GEE-based analyses were used with BMI z-score data from ages 5 and 7 years combined; the β coefficients reflect the average difference in BMI z-score per unit difference in component score. Model 1 adjusts for age (in months) at time of measurement, maternal prepregnancy obesity, birth weight, maternal race/ethnicity, maternal receipt of public assistance during pregnancy, and urinary specific gravity. The DEHP and non-DEHP component score variables are entered into the same regression model and thus the results are mutually adjusted. "Model 2 additionally adjusts for child age 3 urinary metabolite concentration component score. "Model 3 additionally adjusts for child age 5 urinary metabolite concentration DEHP and non-DEHP component scores. "P for interaction with sex. "Interaction test not performed because stratified analyses did not suggest heterogeneity of effect sizes by child sex.

Table 4. Associations^a between maternal prenatal urinary phthalate component scores and child body composition and waist circumference outcomes at age 7 years.

	Percent body fat			Fat mass index			Waist circumference		
Phthalate component	Girls ($n = 164$) β (95% CI) p-value	Boys ($n = 142$) β (95% CI) p-value	<i>p</i> -int ^b	Girls ($n = 164$) β (95% CI) p-value	Boys ($n = 142$) β (95% CI) p-value	<i>p</i> -int ^b	Girls (n = 154) β (95% CI) p-value	Boys (n = 124) β (95% CI) p-value	<i>p</i> -int ^b
Prenatal DEHP component score	-0.14 (-1.19, 0.92) 0.80	-0.31 (-1.56, 0.94) 0.63	ND ^c	-0.01 (-0.42, 0.39) 0.96	-0.11 (-0.56, 0.34) 0.63	ND ^c	-0.11 (-1.46, 1.25) 0.88	-0.65 (-2.26, 0.97) 0.43	NDc
Prenatal non-DEHP component score	0.62 (-0.64, 1.88) 0.33	-1.62 (-2.91, -0.34) 0.01	0.002	0.34 (-0.15, 0.82) 0.17	-0.50 (-0.96, -0.04) 0.03	0.003	1.20 (-0.43, 2.84) 0.15	-2.02 (-3.71, -0.32) 0.02	0.0002

Abbreviations: DEHP, di(2-ethylhexyl) phthalate; ND, not done.

^aAll analyses adjust for age (in months) at time of measurement, maternal prepregnancy obesity, birth weight, maternal race/ethnicity, receipt of public assistance during pregnancy, and urinary specific gravity. The DEHP and non-DEHP component score variables are entered into the same regression model and thus the results are mutually adjusted. The β coefficient reflects the average difference in the outcome variable per unit difference in component score. ^bp for interaction with sex. ^cInteraction test not performed because stratified analyses did not suggest heterogeneity of effect sizes by child sex.

anthropometric outcomes among boys was unexpected; given experimental evidence that phthalates have endocrine-disrupting anti-androgenic or estrogenic effects, we expected a positive, not inverse, association with increased fat mass in boys. Nonetheless, the results do support the hypothesis that phthalates can interfere with the processes of adipogenesis and fat metabolism. We caution against interpreting that phthalates are beneficial in reducing obesity risk, however. Continued ongoing cohort follow-up will reveal whether this inverse association persists, and further analyses will help determine the clinical significance of the association.

The longitudinal cohort design of this study is a major strength because prior crosssectional designs may be biased by diet. Higher food intake leads to both increased body size as well as higher exposures to phthalate contaminants in food, making it difficult to infer causality in cross-sectional analyses (Boas et al. 2010; Colacino et al. 2010; Hatch et al. 2008; Serrano et al. 2014; Stahlhut et al. 2007; Trasande et al. 2013b). Teitelbaum et al. (2012) conducted the only other published prospective study, with urinary phthalate metabolites measured at mean age 7.3 years and BMI data collected on average 1 year later. Among overweight children, higher MEP and molar sum of MEP, MBP, and MiBP were associated with higher BMI; however, no significant associations were observed overall (Teitelbaum et al. 2012). In contrast, we did not observe any statistically significant positive associations between BMI z-score and maternal or child urine phthalate metabolite concentrations.

PCA has been used to assess patterns of exposure from complex chemical mixtures, and may further be particularly advantageous for studying urinary metabolite concentrations (Billionnet et al. 2012; Burstyn 2004; Dominici et al. 2010; Pan et al. 2007). Correlations between metabolite concentrations are expected to reflect exposures to a single parent compound, exposures to several parent compounds with similar exposure routes, and/or the result of parent compound metabolism by the same or similarly induced metabolic pathways; these scenarios are aligned with PCA approaches to construct latent variables reflecting patterns of metabolite concentrations. This was clearly represented by identification of a primary component loading highly for DEHP metabolites, almost certainly reflecting DEHP exposure. Additionally, PCA was useful for identifying a second component that loaded heavily for the non-DEHP metabolites and predicted anthropometric outcomes, whereas analyses of each of the non-DEHP metabolites separately produced results that were generally of borderline statistical significance. A limitation of PCA is that characterization of the components is dependent on the specific study sample data, and observed component scores may not be widely generalizable. However, we found that when applied to NHANES data, PCA found very similar components and component loadings as observed in CCCEH mothers (see Supplemental Material, Tables S1–S3).

Some limitations of this study should be considered. To our knowledge, this is the first longitudinal birth cohort study examining associations between prenatal phthalate exposure and subsequent childhood anthropometric outcomes, and these results must be confirmed in other populations and cohorts. Due to the exploratory nature of this initial work, p-values were not corrected for multiple comparisons, and the significance of the findings may be overestimated. Another limitation was availability of only one spot urine sample per mother, and due to intraindividual variation in urinary metabolite concentrations, exposure was likely categorized with some error (Braun et al. 2012; Frederiksen et al. 2013; Preau et al. 2010; Teitelbaum et al. 2008). Additionally, this cohort consists exclusively of urban African-American and Dominican families, almost all from low-income backgrounds, which limits the generalizability of the work. Further, data on physical activity, diet, and other potential confounders were not available for these subjects; thus bias from residual confounding may exist. However, although diet and physical activity most certainly affect anthropometric outcomes, it is difficult to accurately measure these factors in young children (Bell et al. 2013; Burrows et al. 2010; Chinapaw et al. 2010; de Lauzon-Guillain et al. 2012). Last, the sample size available for sex stratified analyses was limited.

Conclusions

Although cross-sectional studies suggest that phthalate exposures may increase obesity predisposition, our longitudinal birth cohort study did not find any statistically significant positive association between child anthropometric outcomes and maternal prenatal urine phthalate metabolite concentrations. Conversely, there were inverse associations

Table 5. Associations^a between maternal urinary phthalate metabolites and child anthropometric outcomes.

	BMI z-score at	BMI z-score at ages 5 and 7^b		fat at age 7	FMI at	t age 7	Waist circumference at age 7		
Maternal phthalate urinary metabolite	Girls (n = 181)	Boys (n = 156)	Girls ($n = 164$)	Boys (n = 142)	Girls ($n = 164$)	Boys (n = 142)	Girls ($n = 154$)	Boys ($n = 124$)	
	β (95% CI)	β (95% CI)	β (95% CI)	β (95% CI)	β (95% CI)	β (95% CI)	β (95% CI)	β (95% CI)	
	p-value	p-value	p-value	p-value	p-value	p-value	p-value	p-value	
Molar sum of DEHP metabolites	-0.08	-0.09	-0.13	-0.39	-0.03	-0.13	-0.13	-0.65	
	(-0.25, 0.09)	(-0.28, 0.11)	(-1.09, 0.84)	(-1.57, 0.79)	(-0.40, 0.34)	(-0.55, 0.29)	(-1.37, 1.12)	(-2.16, 0.87)	
	0.37	0.39	0.80	0.51	0.88	0.54	0.84	0.40	
MCPP	0.10	-0.24	0.08	-0.63	0.14	-0.19	0.70	-0.88	
	(-0.17, 0.37)	(-0.47, -0.01)	(-1.07, 1.23)	(-1.86, 0.60)	(-0.30, 0.58)	(-0.63, 0.25)	(-0.76, 2.16)	(-2.51, 0.76)	
	0.46	0.04	0.89	0.32	0.52	0.40	0.34	0.29	
MiBP	0.02	-0.17	0.78	-0.97	0.35	-0.29	1.12	-1.18	
	(-0.18, 0.22)	(-0.36, 0.01)	(-0.33, 1.88)	(-2.10, 0.17)	(-0.08, 0.77)	(-0.70, 0.11)	(-0.34, 2.58)	(-2.66, 0.31)	
	0.86	0.06	0.17	0.09	0.11	0.16	0.13	0.12	
MBP	0.11	-0.22	-0.52	-1.05	0.32	-0.34	0.85	-1.34	
	(-0.12, 0.34)	(-0.44, 0.00)	(-0.72, 1.76)	(-2.26, 0.15)	(-0.15, 0.80)	(-0.78, 0.09)	(-0.76, 2.47)	(-2.91, 0.23)	
	0.33	0.05	0.41	0.09	0.18	0.12	0.30	0.10	
MBzP	-0.04	-0.16	0.03	-0.75	0.03	-0.18	0.10	-0.70	
	(-0.18, 0.10)	(-0.31, -0.01)	(-0.81, 0.87)	(-1.64, 0.15)	(-0.30, 0.35)	(-0.50, 0.14)	(-1.03, 1.19)	(-1.87, 0.48)	
	0.56	0.04	0.95	0.10	0.87	0.26	0.86	0.24	
MEP	0.02	-0.04	0.33	-0.86	0.12	-0.32	0.63	-1.27	
	(-0.15, 0.19)	(-0.19, 0.10)	(-0.61, 1.27)	(-1.78, 0.07)	(-0.24, 0.49)	(-0.65, 0.01)	(-0.56, 1.83)	(-2.46, -0.08)	
	0.83	0.56	0.49	0.07	0.50	0.06	0.30	0.04	

Abbreviations: BMI, body mass index; FMI, fat mass index; MBP, mono-*n*-butyl phthalate; MBzP, monobenzyl phthalate; MCPP, mono(3-carboxypropyl) phthalate; MEP, monoethyl phthalate; MiBP, mono-isobutyl phthalate.

All analyses adjust for age (in months) at time of measurement, maternal prepregnancy obesity, birth weight, maternal race/ethnicity, receipt of public assistance during pregnancy, and urinary specific gravity. Metabolite concentrations were In-transformed for analyses. The β coefficient reflects the average difference to in the outcome variable per unit difference in the In-transformed metabolite concentration. ^bGEE-based analyses were used with BMI z-score data from ages 5 and 7 years combined.

between maternal non-DEHP component score and BMI z-score, body fat percentage, FMI, and waist circumference in boys only. However, phthalate exposures have been linked to several poor health outcomes in children (Sathyanarayana 2008; Whyatt et al. 2014), and we caution against interpreting these results to suggest that phthalates are beneficial in reducing obesity risk. Further cohort follow-up will delineate whether negative associations between prenatal exposures and body size in boys persist into adolescence, or if new associations emerge.

REFERENCES

- Arif AA, Shah SM. 2007. Association between personal exposure to volatile organic compounds and asthma among US adult population. Int Arch Occup Environ Health 80:711–719.
- Bell LK, Golley RK, Magarey AM. 2013. Short tools to assess young children's dietary intake: a systematic review focusing on application to dietary index research. J Obes 2013:709626; doi:10.1155/2013/709626.
- Bengraïne K, Marhaba T. 2003. Using principal component analysis to monitor spatial and temporal changes in water quality. J Hazard Mater 100:179–195.
- Bergh C, Torgrip R, Emenius G, Ostman C. 2011. Organophosphate and phthalate esters in air and settled dust – a multi-location indoor study. Indoor Air 21:67–76.
- Biemann R, Navarrete Santos A, Navarrete Santos A, Riemann D, Knelangen J, Blüher M, et al. 2012. Endocrine disrupting chemicals affect the adipogenic differentiation of mesenchymal stem cells in distinct ontogenetic windows. Biochem Biophys Res Commun 417:747–752.
- Billionnet C, Sherrill D, Annesi-Maesano I, GERIE study. 2012. Estimating the health effects of exposure to multi-pollutant mixture. Ann Epidemiol 22:126–141.
- Boas M, Frederiksen H, Feldt-Rasmussen U, Skakkebæk NE, Hegedüs L, Hilsted L, et al. 2010. Childhood exposure to phthalates: associations with thyroid function, insulin-like growth factor I, and growth. Environ Health Perspect 118:1458–1464; doi:10.1289/ehp.0901331.
- Boberg J, Metzdorff S, Wortziger R, Axelstad M, Brokken L, Vinggaard AM, et al. 2008. Impact of diisobutyl phthalate and other PPAR agonists on steroidogenesis and plasma insulin and leptin levels in fetal rats. Toxicology 250:75–81.
- Braun JM, Smith KW, Williams PL, Calafat AM, Berry K, Ehrlich S, et al. 2012. Variability of urinary phthalate metabolite and bisphenol A concentrations before and during pregnancy. Environ Health Perspect 120:739–745; doi:10.1289/ehp.1104139.
- Burrows TL, Martin RJ, Collins CE. 2010. A systematic review of the validity of dietary assessment methods in children when compared with the method of doubly labeled water. J Am Diet Assoc 110:1501–1510.
- Burstyn I. 2004. Principal component analysis is a powerful instrument in occupational hygiene inquiries. Ann Occup Hyg 48:655–661.
- Casals-Casas C, Desvergne B. 2011. Endocrine disruptors: from endocrine to metabolic disruption. Annu Rev Physiol 73:135–162.
- Casals-Casas C, Feige JN, Desvergne B. 2008. Interference of pollutants with PPARs: endocrine

- disruption meets metabolism. Int J Obes (Lond) 32(suppl 6):S53–S61.
- CDC (Centers for Disease Control and Prevention). 2004. A SAS Program for the CDC Growth Charts. Available: http://www.cdc.gov/nccdphp/dnpao/ growthcharts/resources/sas.htm [accessed 20 May 2015].
- CDC. 2014. Fourth National Report on Human Exposure to Environmental Chemicals, Updated Tables, August, 2014. Available: http://www.cdc.gov/exposurereport/pdf/fourthreport_updatedtables_aug2014.pdf [accessed 17 December 2014].
- Chinapaw MJ, Mokkink LB, van Poppel MN, van Mechelen W, Terwee CB. 2010. Physical activity questionnaires for youth: a systematic review of measurement properties. Sports Med 40:539–563.
- Colacino JA, Harris TR, Schecter A. 2010. Dietary intake is associated with phthalate body burden in a nationally representative sample. Environ Health Perspect 118:998–1003; doi:10.1289/ehp.0901712.
- Curtis LH, Hammill BG, Eisenstein EL, Kramer JM, Anstrom KJ. 2007. Using inverse probabilityweighted estimators in comparative effectiveness analyses with observational databases. Med Care 45(10 suppl 2):S103–S107.
- de Lauzon-Guillain B, Oliveira A, Charles MA, Grammatikaki E, Jones L, Rigal N, et al. 2012. A review of methods to assess parental feeding practices and preschool children's eating behavior: the need for further development of tools. J Acad Nutr Diet 112:1578–1602.
- Dominici F, Peng RD, Barr CD, Bell ML. 2010. Protecting human health from air pollution: shifting from a single-pollutant to a multipollutant approach. Epidemiology 21:187–194.
- Frederiksen H, Kranich SK, Jorgensen N, Taboureau O, Petersen JH, Andersson AM. 2013. Temporal variability in urinary phthalate metabolite excretion based on spot, morning, and 24-h urine samples: considerations for epidemiological studies. Environ Sci Technol 47:958–967.
- Grün F, Blumberg B. 2009. Minireview: the case for obesogens. Mol Endocrinol 23:1127–1134.
- Guo H, Wang T, Louie PK. 2004. Source apportionment of ambient non-methane hydrocarbons in Hong Kong: application of a principal component analysis/absolute principal component scores (PCA/APCS) receptor model. Environ Pollut 129:489–498.
- Hao C, Cheng X, Xia H, Ma X. 2012. The endocrine disruptor mono-(2-ethylhexyl) phthalate promotes adipocyte differentiation and induces obesity in mice. Biosci Rep 32:619–629.
- Harris CA, Henttu P, Parker MG, Sumpter JP. 1997. The estrogenic activity of phthalate esters *in vitro*. Environ Health Perspect 105:802–811.
- Hastie T, Tibshirani R, Friedman J. 2009. Principal components, curves and surfaces. In: The Elements of Statistical Learning: Data Mining, Inference and Prediction. 2nd ed. New York: Springer, 534–552.
- Hatch EE, Nelson JW, Qureshi MM, Weinberg J, Moore LL, Singer M, et al. 2008. Association of urinary phthalate metabolite concentrations with body mass index and waist circumference: a cross-sectional study of NHANES data, 1999–2002. Environ Health 7:27; doi:10.1186/1476-069X-7-27.
- Heindel JJ. 2003. Endocrine disruptors and the obesity epidemic. Toxicol Sci 76:247–249.
- Hernán MA, Hernández-Díaz S, Robins JM. 2004. A structural approach to selection bias. Epidemiology 15:615–625.
- Hubbard AE, Ahern J, Fleischer NL, Van der Laan M, Lippman SA, Jewell N, et al. 2010. To GEE or not to GEE: comparing population average and

- mixed models for estimating the associations between neighborhood risk factors and health. Epidemiology 21:467–474.
- Jobling S, Reynolds T, White R, Parker MG, Sumpter JP. 1995. A variety of environmentally persistent chemicals, including some phthalate plasticizers, are weakly estrogenic. Environ Health Perspect 103:582–587.
- Kleinbaum DG, Kupper LL, Muller KE. 1988. Variable reduction and factor analysis. In: Applied Regression Analysis and Other Multivariable Methods. 2nd ed. Boston, MA:PWS-Kent, 595–640.
- Lampa E, Lind L, Hermansson AB, Salihovic S, van Bavel B, Lind PM. 2012. An investigation of the co-variation in circulating levels of a large number of environmental contaminants. J Expo Sci Environ Epidemiol 22:476–482.
- Lovekamp TN, Davis BJ. 2001. Mono-(2-ethylhexyl) phthalate suppresses aromatase transcript levels and estradiol production in cultured rat granulosa cells. Toxicol Appl Pharmacol 172:217–224.
- Newbold RR, Padilla-Banks E, Jefferson WN, Heindel JJ. 2008. Effects of endocrine disruptors on obesity. Int J Androl 31:201–208.
- Pan Z, Gu H, Talaty N, Chen H, Shanaiah N, Hainline BE, et al. 2007. Principal component analysis of urine metabolites detected by NMR and DESI-MS in patients with inborn errors of metabolism. Anal Bioanal Chem 387:539–549.
- Perera FP, Rauh V, Tsai WY, Kinney P, Camann D, Barr D, et al. 2003. Effects of transplacental exposure to environmental pollutants on birth outcomes in a multiethnic population. Environ Health Perspect 111:201–205; doi:10.1289/ehp.5742.
- Preau JL Jr, Wong LY, Silva MJ, Needham LL, Calafat AM. 2010. Variability over 1 week in the urinary concentrations of metabolites of diethyl phthalate and di(2-ethylhexyl) phthalate among eight adults: an observational study. Environ Health Perspect 118:1748–1754; doi:10.1289/ehp.1002231.
- Rengarajan S, Parthasarathy C, Anitha M, Balasubramanian K. 2007. Diethylhexyl phthalate impairs insulin binding and glucose oxidation in Chang liver cells. Toxicol in Vitro 21:99–102.
- Rundle A, Hoepner L, Hassoun A, Oberfield S, Freyer G, Holmes D, et al. 2012. Association of childhood obesity with maternal exposure to ambient air polycyclic aromatic hydrocarbons during pregnancy. Am J Epidemiol 175:1163–1172.
- Sathyanarayana S. 2008. Phthalates and children's health. Curr Probl Pediatr Adolesc Health Care 38:34–49.
- Serrano SE, Braun J, Trasande L, Dills R, Sathyanarayana S. 2014. Phthalates and diet: a review of the food monitoring and epidemiology data. Environ Health 13:43; doi:10.1186/1476-069X-13-43.
- Silva MJ, Barr DB, Reidy JA, Malek NA, Hodge CC, Caudill SP, et al. 2004. Urinary levels of seven phthalate metabolites in the U.S. Population from the National Health and Nutrition Examination Survey (NHANES) 1999–2000. Environ Health Perspect 112:331–338; doi:10.1289/ehp.6723.
- Silva MJ, Samandar E, Preau JL Jr, Reidy JA, Needham LL, Calafat AM. 2007. Quantification of 22 phthalate metabolites in human urine. J Chromatogr B Analyt Technol Biomed Life Sci 860:106–112.
- Stahlhut RW, van Wijngaarden E, Dye TD, Cook S, Swan SH. 2007. Concentrations of urinary phthalate metabolites are associated with increased waist circumference and insulin resistance in adult U.S. males. Environ Health Perspect 115:876–882; doi:10.1289/ehp.9882.

- Takeuchi S, lida M, Kobayashi S, Jin K, Matsuda T, Kojima H. 2005. Differential effects of phthalate esters on transcriptional activities via human estrogen receptors α and β , and androgen receptor. Toxicology 210:223–233.
- Teitelbaum SL, Britton JA, Calafat AM, Ye X, Silva MJ, Reidy JA, et al. 2008. Temporal variability in urinary concentrations of phthalate metabolites, phytoestrogens and phenols among minority children in the United States. Environ Res 106:257–269.
- Teitelbaum SL, Mervish N, Moshier EL, Vangeepuram N, Galvez MP, Calafat AM, et al. 2012. Associations
- between phthalate metabolite urinary concentrations and body size measures in New York City children. Environ Res 112:186–193.
- Trasande L, Attina TM, Sathyanarayana S, Spanier AJ, Blustein J. 2013a. Race/ethnicity-specific associations of urinary phthalates with childhood body mass in a nationally representative sample. Environ Health Perspect 121:501–506; doi:10.1289/ehp.1205526.
- Trasande L, Sathyanarayana S, Jo Messito M, Gross RS, Attina TM, Mendelsohn AL. 2013b. Phthalates and the diets of US children and adolescents. Environ Res 126:84–90.
- Whyatt RM, Perzanowski MS, Just AC, Rundle AG, Donohue KM, Calafat AM, et al. 2014. Asthma in inner-city children at 5–11 years of age and prenatal exposure to phthalates: the Columbia Center for Children's Environmental Health cohort. Environ Health Perspect 122:1141–1146; doi:10.1289/ehp.1307670.
- Zitko V. 1994. Principal component analysis in the evaluation of environmental data. Mar Pollut Bull 28:718–722.